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Electronic Supplementary Material

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**Title: Challenges and Opportunities for Managing Aquatic Mercury Pollution in Altered Landscapes**

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## Supporting Information for Figure 1

The data points presented in Figure 1 are based on a review of the literature reporting measurements of sediment total Hg and MeHg contents at locations impacted by different landscape perturbations (mining, industrial activity, rice production, and urbanization). Only a very limited number of sediment total Hg and MeHg data were available from streams impacted by forestry operations; thus, these data were not included in the figure.

In papers where multiple data points were presented from a similar area (e.g., data from multiple years/seasons and/or to identify within-site spatial variability), we calculated a median value and included this value to represent that specific site in Figure 1. Some papers reported only mean values, and this value was plotted in Figure 1.

## Urban Areas

The urban data (n=121) presented in Figure 1 are from following papers:

Chalmers, A.T., Krabbenhoft, D.P., Van Metre, P.C., Nilles, M.A., 2014. Effects of urbanization on mercury deposition and New England. *Environmental Pollution* 192, 104-112.

Deonarine, A., Hsu-Kim, H., Zhang, T., Cai, Y., Richardson, C.J., 2015. Legacy source of mercury in an urban stream-wetland ecosystem in central North Carolina, USA. *Chemosphere* 138, 960-965.

\* Fleck, J.A., Marvin-DiPasquale, M., Eagles-Smith, C.A., Ackerman, J.T., Lutz, M.A., Tate, M., Alpers, C.N., Hall, B.D., Krabbenhoft, D.P., Eckley, C.S., 2016. Mercury and methylmercury in aquatic sediment across western North America. *Science of the Total Environment* 568, 727-738.

\* Includes data sourced from multiple other studies

Heyes, A., Miller, C., Mason, R.P., 2004. Mercury and methylmercury in Hudson River sediment: impact of tidal resuspension on partitioning and methylation. *Marine Chemistry* 90, 75-89.

Kim, E., Noh, S., Lee, Y.G., Kundu, S.R., Lee, B.G., Park, K., Han, S., 2014. Mercury and methylmercury flux estimation and sediment distribution in an industrialized urban bay. *Marine Chemistry* 158, 59-68.

Marvin-DiPasquale, M., Lutz, M.A., Brigham, M.E., Krabbenhoft, D.P., Aiken, G.R., Orem, W.H., Hall, B.D., 2009. Mercury Cycling in Stream Ecosystems. 2. Benthic Methylmercury Production and Bed Sediment-Pore Water Partitioning. *Environmental Science & Technology* 43, 2726-2732.

Scudder, B.C., Chasar, L.C., Wentz, D.A., Bauch, N.J., Brigham, M.E., Moran, P.W., and Krabbenhoft, D.P., 2009. Mercury in fish, bed sediment, and water from streams across the United States, 1998–2005, U.S. Geological Survey Scientific Investigations Report USGS, p. 74.

Sinclair, K.A., Xie, Q., Mitchell, C.P.J., 2012. Methylmercury in water, sediment, and invertebrates in created wetlands of Rouge Park, Toronto, Canada. *Environmental Pollution* 171, 207-215.

Strickman, R.J., Mitchell, C.P.J., 2017. Methylmercury production and accumulation in urban stormwater ponds and habitat wetlands. *Environmental Pollution* 221, 326-334.

Yang, J., Chen, L., Shi, W.L., Liu, L.Z., Li, Y., Meng, X.Z., 2015. Mercury distribution in sediment along urban-rural gradient around Shanghai (China): implication for pollution history. *Environmental Science and Pollution Research* 22, 1697-1704.

## **Industrial Areas**

The industrial area data (n=33) in Figure 1 are from the following papers:

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- Birkett, J.W., Lester, J.N., 2005. Distribution of mercury and methylmercury in the sediments of a lowland river system. *Proceedings of the Royal Society a-Mathematical Physical and Engineering Sciences* 461, 1335-1355.
- Bloom, N.S., Moretto, L.M., Ugo, P., 2004. A comparison of the speciation and fate of mercury in two contaminated coastal marine ecosystems: The Venice Lagoon (Italy) and Lavaca Bay (Texas). *Limnology and Oceanography* 49, 367-375.
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- Janssen, S.E., Johnson, M.W., Blum, J.D., Barkay, T., Reinfelder, J.R., 2015. Separation of monomethylmercury from estuarine sediments for mercury isotope analysis. *Chemical Geology* 411, 19-25.
- Kim, E., Noh, S., Lee, Y.G., Kundu, S.R., Lee, B.G., Park, K., Han, S., 2014. Mercury and methylmercury flux estimation and sediment distribution in an industrialized urban bay. *Marine Chemistry* 158, 59-68.

- Liu, B., Yan, H.Y., Wang, C.P., Li, Q.H., Guedron, S., Spangenberg, J.E., Feng, X.B., Dominik, J., 2012. Insights into low fish mercury bioaccumulation in a mercury-contaminated reservoir, Guizhou, China. *Environmental Pollution* 160, 109-117.
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### **Rice/Agricultural Areas**

The rice/agricultural data (n=6) presented in Figure 1 are from following papers:

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## Mining Impacted Areas

The mining-impacted area data (n=89) presented in Figure 1 are from following papers:

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United States, 1998–2005, U.S. Geological Survey Scientific Investigations Report USGS, p. 74.

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## Reservoirs

The reservoir data (n=23) presented in Figure 1 are from following papers:

- Becker, J.C., Groeger, A.W., Nowlin, W.H., Chumchal, M.M., Hahn, D., 2011. Spatial variability in the speciation and bioaccumulation of mercury in an arid subtropical reservoir ecosystem. *Environmental Toxicology and Chemistry* 30, 2300-2311.
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### Supporting Information for Figure 3

**Table S1.** Ratio of estimated Hg mass inputs from terrestrial sources (e.g., river, runoff, groundwater intrusion) relative to direct atmospheric deposition to surface water for a variety of aquatic systems. Ecosystems with large terrestrial Hg:atmospheric Hg input ratios are expected to respond more slowly to anticipated reductions in global Hg emissions, while ecosystems with low terrestrial Hg:atmospheric Hg input ratios that are expected to respond more quickly. Modified from (Jonsson 2013).

<b>System</b>	<b>Terrestrial Hg: Atmospheric Hg Input Ratio</b>	<b>Reference</b>
<b>Ocean</b>		
Global ocean (preindustrial)	0.030	Mason and Sheu 2002
Global ocean (current)	0.065, 0.10	Mason and Sheu 2002, Amos et al. 2013
Baltic Sea	0.53	Soerensen et al. 2016
Mediterranean Sea	0.70	Rajar et al. 2007
Arctic Ocean	2.7	Soerensen et al. 2016
<b>Estuaries</b>		
Global estuaries	1	Rolfhus and Fitzgerald 1995
San Francisco Bay (USA)	1.5	MacLeod et al. 2005
Tokyo Bay (Japan)	1.9	Sakata et al. 2006
Long Island Sound (USA)	7.9	Balcom et al. 2004
NY/NJ Harbor Estuary (USA)	80	Balcom et al. 2008
Chesapeake Bay (USA)	1.6	Mason et al. 1999
<b>Lakes and Reservoirs</b>		
Lake Superior (N. America)	0.63	Qureshi et al. 2009
Lake Michigan (USA)	0.31	Qureshi et al. 2009
Little Rock Lake (USA)	0.06	Watras et al. 1996, Qureshi et al. 2009
Lake Champlain (N. America)	1.6	Gao et al. 2006
Big Dam West (Canada)	11	Ethier et al. 2008
Spring Lake (USA)	0.088	Hines and Brezonik 2007
Lake Melville (Canada)	13	Schartup et al. 2015
Dongfeng Reservoir (China)	68	Feng et al. 2009
Wujiangdu Reservoir (China)	27	Feng et al. 2009
Petit-Saut Reservoir (French Guiana)	17	Muresan et al. 2007
Onondaga Lake (USA; historical industrial point source)	310	Qureshi et al. 2009
Clear Lake (USA; historical mining point source)	67 – 586	Suchanek et al. 2009



#### **Explanatory Information for Figure 4**

The runoff from forested catchments was estimated based on the average runoff ratio from 11 catchments described in the following papers: Mason et al. (1997), Balogh et al. (2005), Shanley et al. (2008), Domagalski et al. (2016). Surface emissions from forested catchments were estimated for each of the catchments where runoff data were available and varied depending on the forest type (marine West Coast coniferous, northwest mountains coniferous, or Great Lakes region deciduous), as described in Denkenberger et al. (2012), Eckley et al. (2016). Runoff from logged/harvested catchments was estimated by determining the increase in runoff Hg loads from nine harvested catchments relative to proximate reference/forested catchments described in Porvari et al. (2003), Allan (2009), Sorensen et al. (2009), de Wit et al. (2014), Eklof et al. (2014), Kronberg et al. (2016). The average increase in response to forest harvesting was multiplied by the runoff yield from forested sites described above. Surface emissions from harvested catchments were based on averaged data described in Eckley et al. (2016), Mazur et al. (2014), Eagles-Smith et al. (2016).

Runoff from urban catchments was estimated based on the average runoff ratio from nine urban catchments described in the following papers: Lawson et al. (2001), Eckley and Branfireun (2008), Naik and Hammerschmidt (2011), Domagalski et al. (2016). Emissions from urban surfaces were based on data obtained from six urban sites that are summarized in Denkenberger et al. (2012).

Runoff from catchments containing mining operations were based on the average from eight catchments that contained abandoned mines as summarized in Domagalski et al. (2016). Surface emissions from mine sites were based on the average data from five active and abandoned mining areas described in Engle et al. (2001), Nacht et al. (2004), Eckley et al. (2011), Kocman and Horvat (2011).

All runoff and emission values were scaled to a common deposition rate of  $10 \mu\text{g m}^{-2}$  value to allow comparisons of the effect of different land-use perturbations between studies. The net landscape sink or source values were based on the difference between Hg inputs from deposition and Hg outputs from runoff and surface emissions.

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